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2 **FOREST VALUE INFERENCE USING CONTINGENT VALUATION AND CHOICE**
3 **EXPERIMENTS**

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1 **Abstract**

2
3 A way to avoid unnecessary costs in estimating the social value of forests is to base such
4 estimations in other values already obtained. This practice has advantages and disadvantages.
5 Among the latter, the reliability of the obtained value is probably the one of most concern. This
6 paper tests the reliability of a specific value inference exercise for forests, finding positive results
7 when contingent valuation and choice experiments are used.

8
9 Key words: Value inference; Choice experiment; Contingent valuation; Forest externalities;
10 Social value; Total economic value.

11

12

13 **Introduction**

14

15 In order to make sound decisions for the whole society, forest planning and management ought to
16 take into account the value of forests for both the landowner and the rest of effected persons. The
17 field of economics helps in this process by being able to estimate the value, in monetary units, of
18 the forest at stake. While the value for the owner –or *private value*– is expressed in terms of
19 market prices, tools to estimate the forest value to the whole society –often called *social value* or
20 *total economic value*– have been developed in the last few decades. The most used one is the
21 contingent valuation method (CVM), with an increasing popularity of other stated preference
22 methods, like choice experiments (CE).

23

1 Unfortunately, these methods are considerably time-consuming and rather expensive to
2 undertake, especially if they are to be applied to a large number of forests. A foreseeable
3 consequence of increasing costs is a decrease in the likelihood to undertake those valuation
4 studies. To some extent, this has been mitigated in areas like transportation or health by
5 “borrowing” or “transferring” some already obtained values to estimate the value of a relatively
6 similar good (Ben-Akiva, and Bolduc, 1987; Koppelman *et al.* 1985; Galbraith and Hensher,
7 1982). However, relatively less attention has been paid to this approach in environmental
8 economics, and even less in forest economics (Xu and Adamowicz, 1997; Azqueta and Touza-
9 Montero, 2000; Rosenberger and Loomis, 2001; Haener *et al.*, 2001). The strongest concern in
10 these fields has probably been the reliability of the procedure, whether the new values obtained
11 from existing ones are similar or significantly different than the value that would be obtained
12 from an *ad hoc* exercise (Loomis, 1992; Bergland *et al.*, 1995; Kirchhoff *et al.*, 1997; Morrison
13 and Bennett, 2000; Morrison *et al.*, 2002).

14
15 This paper proposes a way to infer values from existing ones and conducts a test of predictability
16 of forest social value inference using two different methods, CVM and CE. This is related to
17 benefit transfer tests, although it is different in the sense that there is no transfer between sites but
18 between goods and attributes in the same site. To our knowledge, no similar tests have been
19 conducted combining CVM and CE. Next section introduces the concept of value inference and
20 the test to be conducted. In section three a brief overview of the methods used and their
21 underpinning theory is presented, followed by the details of the exercises designed to apply the
22 inference test. Section four presents the joint preference analysis. Before conclusions, the results
23 of the test are presented.

24

1 **Value Inference and Benefit Transfer**

2

3 The closest literature to the study presented here is the benefit transfer (BT) one. Following
4 Desvougues *et al.* (1992), benefit or value transfer is the use of the monetary value of a good
5 obtained in a given context (often called *study site*), to estimate the value of a similar good under
6 a different context (*policy site*). For instance, the value of a forest can be estimated from the
7 known value of another similar forest placed elsewhere. Sometimes, however, it is the good itself
8 that varies in one way or another. This would be the case of estimating the value of a forest with a
9 given composition from a forest in the same area but with a different composition or management
10 program. This case is what is called here “value inference” to distinguish it from value or benefit
11 transfer.

12

13 Obtaining values from transfer exercises has pros and cons. On the negative side, transferred
14 values are subject to different measurement errors (Bergstrom and De Civita, 1999). This could
15 be specially the case when the study and policy sites differ very much in some characteristics.
16 Potential statistically significant discrepancies between transferred values and values obtained
17 directly have raised concerns about the reliability of transferring benefits (Brouwer, 2000). On
18 the positive side, if reliable, transferability (inference) would be a cost-effective way to estimate
19 the social value of forests. Applying previous research findings might be an attractive alternative
20 to expensive and time consuming original research.

21

22 Transferring is standard practice for some practitioners, especially in cost-benefit analysis
23 (Bergstrom and De Civita, 1999), and it has been suggested that benefit transfer approaches will
24 become even more widespread (Desvougues *et al.*, 1992; OECD, 1993). One reason for this

1 practice is that since the 1990s more rigorous approaches to benefit transfer have been proposed
2 with the use of better protocols, cautions, and common practice recommendations (Desvouges *et*
3 *al.*, 1992; Boyle and Bergstrom, 1992; Smith, 1992; Kask and Shogren, 1994; Brouwer, 2000;
4 EFTEC, 2000; Rosenberger and Loomis, 2000; Ruijgrok, 2001).

5
6 Most of the comments above can be applied to value inference. It shares the basic advantages and
7 disadvantages of BT and different implementation approaches are possible, similar to BT. The
8 rest of this section is devoted to illustrate the way value inference could work in a forest valuation
9 context and to introduce the test to be undertaken in the case study application.

10

11 In the proposed application, two forests are considered, FOREST A and FOREST B, placed in
12 the same region, but with different characteristics (attributes). The idea of the test is to see
13 whether the value of FOREST B can be predicted (inferred) from the value of FOREST A using
14 the marginal values of the attributes that vary from one forest to the other estimated with a choice
15 experiment.

16

17 A bit more formally, assume that the correct welfare measure from a rational individual reflecting
18 her value (maximum willingness to pay) for the provision of a good is the compensating variation
19 (*CV*), which is the amount of money that deducted to her income (*y*) leaves her indifferent
20 between getting the good by paying this amount, and not getting the good and paying nothing.
21 Therefore, for a public good *z* and all attributes *j* of this good,

22

23 [1] $v(p, y, z^0(z_j)) = v(p, y-CV, z^1(z_j))$

24

1 where v is the indirect utility level of the individual, p is the vector of prices of private goods, z^0
 2 is the level of the public good before the proposed change, z^1 is the new level of the good, and z_j
 3 denotes the j relevant attributes of good z .

4

5 The values CV_j assigned to each relevant attribute $j = 1, \dots, J$, of good z , can be reflected in terms
 6 of an indirect welfare function as

7

$$8 \quad [2] \quad v(p, y, z^0(z_1, \dots, z_J)) = v(p, y - CV(CV_1, \dots, CV_J), z^1(z_1, \dots, z_J))$$

9

10 The general expression [2] takes the form

11

$$12 \quad v(p, y, z^0_{1+\dots+J}) = v(p, y - CV_1 - \dots - CV_J, z^1_{1+\dots+J})$$

13

14 when independence between attributes (no higher order cross effects) is assumed and the utility
 15 function is linear and additive with respect to the attributes.

16

17 Similar to [1], the indirect utility for FOREST A and FOREST B can be expressed respectively as

18

$$19 \quad v(p, y, z^0(z_j)) = v(p, y - CV^A, z^A(z_j))$$

$$20 \quad v(p, y, z^0(z_j)) = v(p, y - CV^B, z^B(z_j)),$$

21

22 where z^0 denotes the status quo situation (marginal agricultural use of the land, in the case study
 23 below), z^A and z^B denote the goods FOREST A and FOREST B, and compensating variations

1 CV^A and CV^B reflect the maximum willingness to pay (WTP) of the individual to obtain each
2 good.

3
4 In the application below, CV^A for FOREST A and CV^B for FOREST B are estimated by CVM. In
5 addition, CV^B is also estimated from inference by adding to CV^A the corresponding value of
6 attributes variation between both forests (CV_j). The value of these attributes is estimated by CE.
7 To distinguish between both CV^B , the one corresponding to the mean of the maximum
8 willingness to pay calculated from CVM will be denoted μ_{CVM}^B , while the one obtained by value
9 inference will be denoted as μ_{VI}^B .

10

11 The null and alternative hypothesis for the predictability (or convergence between *ad hoc* and
12 inferred values) test can then be expressed as

13

14 [3] $H_0: \mu_{CVM}^B = \mu_{VI}^B$

15 $H_1: \mu_{CVM}^B \neq \mu_{VI}^B$

16

17 If the null hypothesis cannot be rejected at a given confidence level, the reliability of obtaining
18 values from value inference could not be discarded, when estimated by the methods used in the
19 empirical exercise.

20

21

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1 **Afforestation in Catalonia and the research design**

2

3 **Choice experiment method**

4 *Choice experiment*, also called *contingent choice*, is a stated preference valuation method that is
5 becoming progressively more popular. Other stated preference methods are contingent ranking
6 and contingent rating o paired comparisons –see Louviere *et al.* (2000) or Hanley *et al.* (2001) for
7 an overall description of stated preference methods. The choice experiment approach consists in a
8 set of options, usually called *alternatives*, containing common attributes of a good with different
9 values, often called *levels*, for each attribute. One of the attributes is the money a person would
10 have to pay, or get in compensation, for the overall good as described by its attributes. It could
11 also be applied to a bundle of goods, and the “attributes” would be the different goods of the
12 bundle. Through a questionnaire, a sample of the population is faced with a number of
13 alternatives, out of which a person has to choose the most preferred one. The alternatives
14 presented to respondents ought to include always the status quo situation. The information of
15 individual choices is then used to econometrically estimate the marginal value of each attribute
16 (Hanley *et al.*, 1998a; Hanley *et al.*, 1998b; Morrison *et al.*, 1998; Rolfe *et al.*, 2002).

17

18 Initially, marketing was one of the main fields of application of choice experiments, but other
19 areas, such as transportation, geography, and the environment, among others, incorporated the
20 method (Louviere, 1991). Most of the papers on choice experiments in environmental economics
21 are relatively recent (Opaluch *et al.*, 1993; Adamowicz *et al.*, 1994; Eom, 1994; Adamowicz *et*
22 *al.*, 1996; Rolfe and Bennett, 1996; Boxall *et al.*, 1996; Bergland, 1997; Adamowicz *et al.*, 1998;
23 Hanley *et al.*, 1998a; Hanley *et al.*, 1998b; Boxal and Macnab, 2000; Rolfe *et al.*, 2002).

1
2 The number of papers dealing with BT and choice experiments is rather modest. The closest
3 applications to the one presented here are those of Morrison *et al.* (1998) and Morrison and
4 Bennett (2000). These authors estimated the marginal WTP for some attributes of Australian
5 wetlands, using the choice experiment method in three subsamples, and tested the equality among
6 those marginal WTP values. They concluded that choice experiments estimate similar marginal
7 values when applied to goods with different levels of their attributes.

8
9 The valuation exercise presented in this paper involves the estimation of the impact of alternative
10 afforestation programs on non-market forest values. The afforestation program concerned was in
11 Catalonia, a region in the Northeast of Spain, which has 1.3 million ha of forests, or about 40 per
12 cent of its total area. Although the composition of the forest varies from the coastal areas to the
13 Pyrenees and the inland plains, most are composed of Mediterranean species. The pine is the
14 dominant species covering 50 per cent of the forested area, followed by the holm oak with some
15 10 per cent (Ministerio de Medio Ambiente, 1996). The majority (77 per cent) of forests is
16 privately owned.

17
18 The program being proposed involves an increase in forest coverage from the current 40 per cent
19 of the Catalonian area to 50 per cent. The additional 10 per cent of forest area would be at the
20 expense of marginal agricultural land. Each questionnaire contained a sequence of four election
21 sets with three alternatives. The alternatives varied from questionnaire to questionnaire, as will be
22 explained below. One of the three alternatives was always the status quo situation, i.e. no
23 afforestation and no payment required. The other two reflected 10% afforestation with different
24 attribute levels.

1
 2 The first step in implementing the CE exercise was the determination of the attributes to be used
 3 to describe each afforestation alternative. Discussions with experts working in forestry research,
 4 focus groups and repeated interviews with samples of the Catalan population, were conducted to
 5 determine the non-market attributes associated with the forest. The attributes chosen for the
 6 analysis were some recreational activities –such as picnicking, picking mushrooms, and driving
 7 motor vehicles on forest ways–, CO₂ sequestration, erosion prevention, and the payment
 8 contribution. Payment values were originally expressed in Spanish pesetas, although in this paper
 9 they are reported in euros. The attributes and their levels are shown in Table 1.

10 **Table 1.** Attributes and levels used in the choice experiment

Attribute	Description	Levels
Picnic	Picnicking allowed in the new forests	Yes No
Drive	Driving by car through the new forests would be allowed	Yes No
Mushrooms	Picking mushrooms allowed in the new forests	Yes No
CO ₂	CO ₂ sequestered annually by the new forests. Equivalent to the pollution produced annually by a city of...	300.000 people 400.000 people 500.000 people 600.000 people
Erosion	Erosion decrease (If not afforested, land would become unproductive....)*	After 100 years After 300 years After 500 years After 700 years
Cost ^{###}	The afforestation cost per person and year	6 euros 12 euros 18 euros 24 euros

11 **Note:**

12 [#]SQ: Status Quo

13 ^{###} Payment values were originally expressed in Spanish pesetas, although in this paper they are reported in euros

14 * If afforested, erosion would be prevented indefinitely.

15

1 The final design selected 64 pairs of alternative afforestation compositions out of a universe of
2 $(2^3 \times 4^3) \times (2^3 \times 4^3)$ possible combinations, following a fractional factorial design (Louviere, 1988).
3 The 64 pairs were then blocked into 16 versions of 4 choice sets of two alternatives plus the
4 status quo. In this way, there were a total of 16 questionnaire versions, each one assigned
5 randomly to a subsample out of a larger sample of 1200 individuals representative in terms of
6 location, age and gender of the Catalan population over 18 years old. Appendix A displays one of
7 the resulting choice sets.

8

9 **Contingent valuation method**

10 A second valuation exercise was undertaken, to find the maximum willingness to pay for two
11 given new forest compositions: FOREST A and FOREST B. The new forests had the same
12 extension than in the previous exercise, i.e. covering an additional 10% of Catalonia, substituting
13 marginal agricultural land. The method used was the Contingent Valuation Method, which was
14 applied in a similar way than the choice experiments, but without varying the physical attributes
15 of the forest, only the monetary payments. CVM is a stated preference method where respondents
16 are asked for their maximum willingness to pay (or minimum willingness to accept in
17 compensation) to get or avoid a given increase or decrease in environmental quality (see, for
18 instance, Mitchell and Carson, 1989). CVM has been used to estimate the value of a wide variety
19 of environmental resources. However, its use has been subject to criticism in terms of its ability
20 to deliver reliable and accurate estimates of the willingness to pay and the correct design of CVM
21 surveys (Diamond and Hausman, 1994).

22

23 Following the guidelines of the NOAA blue ribbon panel (Arrow *et al.*, 1993), a binary
24 dichotomous choice format was used where each forest was offered to individuals at a given price

1 (bid). Thus, after describing the particular forest composition, the respondents were given the
 2 option of choosing to pay the bid (annual payment) and accepting the afforestation program, not
 3 paying the cost amount and not accepting the program (“status quo” option) or responding “I
 4 don’t know”. The monetary amounts were 6, 12, 18, 36, 48, and 72 euros, and one amount was
 5 assigned to each questionnaire version, making a total of six versions per type of afforestation.
 6 Table 2 shows the attribute levels of both forests. It is assumed, from the differences in physical
 7 attribute levels that FOREST B is preferred to FOREST A. Appendix B shows a card
 8 corresponding to the FOREST A CVM exercise.

9
 10 **Table 2.** Attributes and levels used in the CVM exercises

Attribute	FOREST A	FOREST B
Picnic	Yes	Yes
Drive	Yes	Yes
Mushrooms	Yes	Yes
CO ₂	400,000 people	600,000 people
Erosion	After 500 years	After 100 years
Cost	6 euros 12 euros 18 euros 36 euros 48 euros 72 euros	6 euros 12 euros 18 euros 36 euros 48 euros 72 euros

11
 12 A total sample of 1000 people, representative in terms of location, age and gender of the Catalan
 13 population over 18 years old, was divided into two roughly equal sub-samples¹. Each sub-sample

¹ The sample of 1000 respondents used in the CVM exercises was a sub-sample of the total sample used in the CE. Each respondent faced a questionnaire that contained a description of the positive and negative effects of the afforestation programme, the choice experiment questions, the dichotomous choice CVM question, a part devoted to debriefing, and a socio-demographic section.

1 was assigned a given afforestation project (FOREST A or FOREST B), and one of the six CVM
2 questionnaire versions were randomly allocated to respondents within each sub-sample.

3
4 **Sample**
5 Interviews were conducted face-to-face in households, in the fall of 1999. The average valid
6 response rate was 95 per cent across all subsamples for the CE application and 84 per cent for the
7 CVM survey. The difference is explained by a higher number of protest answers in the CVM
8 version. The most common motives for protesting were that the “government should pay”, and
9 the belief that the provision of the good would be ineffectual. Protest answers were excluded for
10 the value estimation. The socio-demographics of the respondents can be seen in detail in Mogas
11 *et al.* (2002).

12
13 **Model specification**
14 The CE structure and the referendum CVM structure can both be analyzed using a random utility
15 model (RUM) (Thurstone, 1927; McFadden, 1974). Under the RUM framework, the election of
16 one of the three scenarios in the choice experiment, or yes/no in the CVM experiment, represents
17 a discrete choice from a set of alternatives reflecting an underlying utility function. The utility
18 that an individual derives from an alternative i is represented as

19
20
$$U_i = v_i + \varepsilon_i$$

21

1 This conditional indirect utility function is comprised of a deterministic component (v_i) and a
 2 random error component (ε_i). An individual will choose alternative i if $U_i > U_j$ for all $j \neq i$. Thus,
 3 the probability of choosing alternative i is

4

$$5 \quad \pi(i) = \text{Probability}(v_i + \varepsilon_i \geq v_j + \varepsilon_j), \quad \forall j \in C$$

6

7 where C is the choice set. If the random terms are assumed to be independently and identically
 8 distributed with an extreme value type I distribution, then the probability of choosing i takes the
 9 form

10

$$11 \quad \pi(i) = \frac{e^{\omega v_i}}{\sum_{j \in C} e^{\omega v_j}}$$

12

13

14 where ω is a scale parameter, inversely proportional to the standard deviation of the error
 15 distribution, and conventionally normalized to one. In the dichotomous choice form of the CVM,
 16 this formulation can be estimated using the binary logit model (Hanemann, 1984), while in the
 17 CE approach this form represents the conditional logit model (McFadden, 1974).

18

19 **Equivalence test of the models**

20 Before proceeding with the test of the hypothesis for the predictability (or convergence of values)
 21 described above and the estimation of welfare measures, the preference equality across the three
 22 sets of data (CE and CVM for FOREST A, and CVM for FOREST B) is considered. Since the
 23 CE and CVM approaches share the random utility maximization theory, and the choices are being
 24 made over the same goods, it is possible to combine the three datasets for a joint estimation to

1 identify the relative scale factors (Swait and Louviere, 1993; Adamowicz *et al.*, 1994;
2 Adamowicz *et al.*, 1998; Rolfe *et al.*, 2002).

3
4 To see whether the preferences reflected in the choice experiments and the contingent valuation
5 exercises are the same, the equivalence test originally proposed by Swait and Louviere (1993)
6 was conducted. The preference equality across data sources involves pooling the three data
7 sources under the restriction that the common parameters across the models are equal, while
8 controlling for the scale parameters. Although it is not possible to identify the three scale
9 parameters, a relative scale parameter can be computed across separate data sets. The scale factor
10 will account for differences in the variation of the unobserved effects, i.e. error variance
11 heterogeneity (Swait and Louviere, 1993). The scale parameter for the CE dataset is set to 1,
12 while the scale parameters for CVM (from FOREST A and FOREST B) are unconstrained.
13 Therefore, two relative scale parameters with respect to the CE data scale could be estimated: the
14 ratio of the scale parameter of the FOREST A dataset relative to the scale parameter of the CE
15 data (ω_A) and the scale factor of FOREST B relative to the CE data (ω_B).

16
17 The hypothesis to be tested is $H_0: \beta = \beta^{CE} = \beta^{CVMA} = \beta^{CVMB}$, where β^{CE} , β^{CVMA} , β^{CVMB} are utility
18 parameters for the common attributes between the three datasets. Under this hypothesis, the
19 relative scale factors between models can be estimated by stacking the three datasets and
20 conducting a one-dimensional search using a range of values for the scale parameters. The value
21 of the scale parameter is found when the log-likelihood of the conditional logit model using the
22 stacked data, is maximized. The joint model is presented in Table 3. The maximum log-

1 likelihood value of the joint or “stacked” model was achieved when ω_A assumed a value of 1.45
 2 and ω_B a value of 1.61.

3

4 **Table 3.** Joint model estimation (scaled)

Variable	Coefficient	Statistic	Value
Constant	1.146***	Scale parameters	
Picnic	0.073**	ω_A	1.45
Mushrooms	0.106***	ω_B	1.61
Drive	-0.166***		
CO ₂	0.730E-06***		
Erosion	-0.264E-03***	Log. Likelihood	-5046.064
Cost	-0.0180E-01***	Pseudo-R ² (McFadden, 1974)	0.15
Age	-0.010***		
Gender	-0.290***	χ^2	1737.72
Income	0.3195E-05**	Observations	5411
Visitation	0.374***		
Rural	0.607***		

5 ***Significant at 1% level; **Significant at 5% level

6 **Variable definitions:**

7 Age = individual’s age minus mean age of sample (45.64)

8 Gender = gender of the respondent (1 for male, 0 for female)

9 Income = income of the respondent in euros

10 Visitation = Use of the forest for recreation (1 if the respondent had used the forest during the last year, 0
 11 otherwise)

12 Rural = village size (1 if <10,000 inhabitants, 0 if >10,000 inhabitants)

13

14 The Swait and Louviere (1993) test is used to determine whether the null hypothesis should be
 15 rejected. This test statistic is asymptotically chi-squared distributed with $k-1$ degrees of freedom,
 16 where k is the number of attributes that are forced to be the same across datasets. The estimated
 17 chi-squared statistic is 4.724 (5 d.f.), and the critical value of the χ^2 statistic at the 5%
 18 significance level is 11.07, implying that the null hypothesis of equality of parameters cannot be
 19 rejected. In other words, the hypothesis of preference equality between the three data sources can
 20 be retained.

1

2 **Results**

3

4 **Choice experiment results**

5 A conditional logit model was specified and estimated from the choice data using a maximum
6 likelihood approach (McFadden, 1974). The results of the logit analysis of the stated preference
7 data are presented in Table 4.

8

9 **Table 4.** conditional logit model estimation for the CE application

Variable	Coefficient	Statistic	Value
Constant	1.0446***	Log. Likelihood	-4515.908
Picnic	0.0809**	Pseudo-R ²	0.10
Mushrooms	0.1079***	(McFadden, 1974)	
Drive	-0.1649***	Pseudo-R ²	0.27
CO ₂	0.8110E-06***	(Veall- Zimmermann, 1996)	
Erosion	-0.3030E-03***		
Cost	-0.0187***	χ^2	1005.60
Age	-0.0105***	Observations	4576
Gender	-0.3712***		
Income	0.840E-03***		
Visitation	0.4174***		
Rural	0.7385***		

10 **Note:** ***Significant at 1% level; **Significant at 5% level

11

12 **Variable definitions:**

13 Age = individual's age minus mean age of sample (45.64)

14 Gender = gender of the respondent (1 for male, 0 for female)

15 Income = income of the respondent in euros

16 Visitation = Use of the forest for recreation (1 if the respondent had used the forest during the last year, 0
17 otherwise)

18 Rural = village size (1 if <10,000 inhabitants, 0 if >10,000 inhabitants)

19

1 The signs of the parameters are consistent with usual expectations, and almost all attributes are
2 statistically significant at 95% confidence level. The intercept term indicates that the respondents
3 perceive as positive the afforestation program. Picnicking, picking mushrooms, and the CO₂
4 sequestered by the new forests are factors that positively affect utility, while respondents place a
5 negative value to allowing the use of cars in forest ways and to increases in erosion if the
6 afforestation would not be undertaken. The coefficient of the annual contribution is also negative,
7 as expected, which indicates that it is perceived as a cost.

8
9 Focusing on the socio-demographic characteristics, being female, living in a rural area, having
10 higher income, and using forest for recreation, increase significantly the probability of choosing
11 the afforestation program alternatives. Also, ages between 25 and 65 tend to be more willing to
12 pay for the program than younger or older population.

13
14 The fit of the model is rather low², although in line with some other CE applications (Adamowicz
15 *et al.*,1998; Christie and Azevedo, 2002; Morrison *et al.* 2002; Rolfe *et al.*, 2002). It would be
16 advisable, however, to replicate the value inference test with higher fit statistics.

17
18 The β coefficients from the conditional logit model were used to estimate the marginal WTP for
19 each attribute, or implicit price. The marginal annual WTP for each attribute can be inferred by
20 calculating the ratio $-\beta_i/\beta_{COST}$, where β_i is the regression coefficient of the physical attribute and
21 β_{COST} represents the monetary attribute coefficient. The confidence intervals for the different
22 marginal WTP estimations were obtained following the Krinsky and Robb (1986) bootstrapping

² Two Pseudo-R² s based on different construction criteria are used following the recommendations of Amemiya (1981)

1 procedure and the percentile method of Efron and Tibshirani (1993) with 1000 random
 2 extractions³. Table 5 shows the main results.

3
 4 **Table 5.** Marginal annual WTP for
 5 the CE attributes (in euros of 1999)

Variable	Marginal WTP
Picnic	4.33 [#] (0.039, 9.96)
Drive	-8.83 (-15.91, -4.18)
Mushrooms	5.78 (1.07, 12.00)
CO ₂	4.345E-05 (2.18E-05, 7.66E-05)
Erosion	0.017 (0.0064, 0.032)

6 **Note:** [#] Non-significant coefficient in the conditional logit estimation.
 7 In brackets, confidence intervals at 95%.
 8

9 Marginal WTP indicate, for each attribute, the average amount of euros that a person would be
 10 prepared to pay annually, indefinitely, for an increase (or decrease) of one unit in the attribute
 11 level. A positive (negative) marginal value for an attribute denotes that the average person would
 12 be better off with an increase (decrease) in the level of the attribute. The values of PICNIC,
 13 MUSHROOMS, and DRIVE correspond to a discrete change, from being able to picnic, pick
 14 mushrooms, or drive cars in the new forests, to not being able to do those recreational activities.
 15 The marginal WTP for CO₂ reflects the value that a new forest provides to society by
 16 sequestering the CO₂ emissions that a citizen of Catalonia generates annually in production and

³ It involves the simulation of an asymptotic distribution of the coefficients that are generated in a CE, from which confidence intervals can then be computed. The distribution is achieved by taking repeated random draws of the coefficient vectors from the multivariate normal distribution defined by the coefficient estimates and their associated covariance matrix. Implicit prices can then be calculated from each of the random draws of coefficients, and confidence intervals estimated by identifying the values at each tail of the distribution of implicit prices (Rolfe *et al.*, 2002)

1 consumption activities⁴. Similarly, the marginal WTP for EROSION is interpreted as the average
 2 maximum WTP for increasing one year the time horizon of land productivity due to the
 3 prevention of erosion with the land use change. The value is taken with a positive sign, although
 4 the sign is negative in Table 4 due to the way the question was framed.

6 **Contingent valuation results**

7 The coefficients of the two CVM exercises were estimated using a logit regression. Table 6
 8 shows the model estimation results for FOREST A and FOREST B, with the dependent variable
 9 reflecting the answers to accept the proposed annual payment (bid).

11 **Table 6.** Logit model estimation for Forests A and B

Variable	FOREST A Coefficient	FOREST B Coefficient
Constant	0.577 ***	0.802**
Cost	-0.033***	-0.018***
Gender	-0.215E-02	-0.458**
Income	4.261E-4***	4.128E-4**
Visitation	0.581***	0.482**
Rural	1.091***	0.405*
Maximum Log. Likelihood	-289.469	-238.325
% of correct predictions	68%	63%
χ^2	64.856	22.586
Number of valid observations	464	371

12 **Note:** ***Significant at 1% level; **Significant at 5% level

13 **Variable definitions:**

14 Age = individual's age minus mean age of sample (45.64)

15 Gender = gender of the respondent (1 for male, 0 for female)

16 Income = income of the respondent in euros

17 Visitation = Use of the forest for recreation (1 if the respondent had used the forest during the last year, 0
 18 otherwise)

19 Rural = village size (1 if <10,000 inhabitants, 0 if >10,000 inhabitants)

⁴ This estimation was based in the total of CO₂ emissions added to the atmosphere in Catalonia in 1995, equivalent to 6.8 tonnes per year (Departament de Medi Ambient de la Generalitat de Catalunya, 1996)

1
 2 As expected, the sign of PRICE coefficient is negative and significant, which indicates that the
 3 probability for people to accept to pay the proposed amount decreases as the demanded payment
 4 increases. In both CVM models those who have more income, those who used the forest for
 5 recreation during the last year and those who live in a rural area, are more likely to pay. The chi-
 6 square statistic indicates that each model is significant at the 99% level.

7
 8 Table 7 shows the estimation of the mean WTP based on the information from Table 6. Since the
 9 assumed distribution for WTP in the logit model is a symmetrical one, the mean and median
 10 WTP coincide and can be estimated from the ratio $-\alpha/\beta$, where β is the value of the coefficient of
 11 the cost variable in the estimated logit equation, and α is the sum of all other terms in the
 12 equation evaluated at the mean values of the explanatory variables (Hanemann, 1984; Hanemann
 13 and Kanninen, 1999). Table 7 reports the mean WTP for both forests, as well as the confidence
 14 intervals. The confidence intervals were calculated using the Krinsky and Robb (1986)
 15 bootstrapping procedure with 1000 extractions.

16

17 **Table 7.** Mean annual WTP for FOREST A and FOREST B (in euros of 1999)

Welfare specification	WTP FOREST A	WTP FOREST B	P-value (Poe <i>et al.</i>)
Mean CVM	37.5 (31.57, 45.97)	60.95 (47.28, 100.62)	0.00

18 *Note:* 95 per cent confidence intervals for 1000 extractions.

19

20 The mean WTP for FOREST B (WTP_B) is greater than the WTP for FOREST A (WTP_A), by a
 21 difference of about 23 euros. The sign of this result is consistent with expectations, since

1 FOREST B type was supposed to be more desirable than FOREST A because it sequesters more
2 CO₂ and prevents erosion for a longer period of time (Table 2).

3
4 To see, however, if the difference is statistically significant, the one-tail non-parametric test
5 proposed by Poe *et al.* (1997) was applied. This test was constructed following a Krinsky and
6 Robb (1986) bootstrap procedure, with a thousand extractions. The 95% confidence intervals and
7 the probability value for the test of equality among means are shown in Table 7. For a 95%
8 confidence level, the null hypothesis of equal WTP for forests A and B is rejected.

10 **Value Inference Test**

11 The value inference test checks whether μ_{CVM}^B is significantly different than μ_{VI}^B . The value of
12 μ_{CVM}^B in [3] is the WTP for FOREST B reported in Table 7 (60.95 euros per year and person, in
13 1999 values). The value of FOREST B inferred from FOREST A, μ_{VI}^B , is calculated from the
14 WTP for FOREST A and the marginal values obtained through the CE of the two attributes that
15 vary between forests (CO₂ and EROSION), as shown in Table 5 (52.70 euros).

16
17 For the inference test both the values for the increase in CO₂ sequestration of FOREST B
18 compared to FOREST A –which is the annual equivalent of a city of 200,000 people–, and the
19 improvement in erosion prevention –for an extra 400-year period–, were treated as if they
20 responded to a linear function. This is a usual hypothesis in the literature and implies the
21 assumption that marginal unitary values would be constant within the limited segments of level
22 variation examined.

23

1 Table 8 shows the above-mentioned results and their confidence intervals using Krinsky and
 2 Robb (1986) bootstrap procedure with 1000 extractions. It also shows the results of the test for
 3 [3], based on the non-parametric indicator proposed by Poe *et al.* (1997), with 1000 bootstrapping
 4 extractions. In this particular contrast, [3] takes the form:

$$5 \quad H_0: \mu_{CVM}^B - \mu_{VI}^B = 0$$

$$6 \quad H_1: \mu_{CVM}^B - \mu_{VI}^B \neq 0.$$

7
 8
 9 **Table 8.** Predictability test for FOREST B (in euros of 1999)

Welfare specification	WTP FOREST B μ_{CVM}^B	WTP FOREST B PREDICTED μ_{VI}^B	P-value (Poe <i>et al.</i>)
Mean CVM	60.95 (47.28, 100.62)	52.70 (46.14, 64.29)	0.20

10
 11 The result implies that the null hypothesis of equality between the transferred predicted value and
 12 the one obtained with an *ad hoc* estimation cannot be rejected at a 95% confidence level (the *p*
 13 value is above the 5% threshold level). These results suggest that the marginal WTP values
 14 obtained from a choice experiment may constitute an acceptable option for estimating the value
 15 of new forests where attributes of the good vary, instead of undertaking successive CVM studies
 16 to find the value of each forest.

17

18 **Conclusions**

19 In their decision-making, public administrations or private agents may often be interested in
 20 knowing the economic value of environmental assets, like forests. This value is most usually
 21 obtained by applying the contingent valuation method (CVM) or some other valuation method. If
 22 the process of estimating such values were less expensive and time consuming, environmental
 23 values would probably influence decisions more often, contributing to their rationality. A

1 possible cost-effective way for obtaining non-market values for a number of environmental assets
2 (like forests) would be through a procedure with similarities to benefit transfer, named here
3 “value inference”, i.e. relying on previously obtained values. In other words, if by value inference
4 the analyst could obtain approximately the same value than with a CVM *ad hoc* study, the
5 procedure would be an advisable cost-effective alternative.

6
7 This paper has proposed the combined use of CVM and choice experiments (CE) to obtain a base
8 value and marginal attribute values respectively to be used in the inference exercise. Furthermore,
9 it has presented an application between forests that vary in some attribute levels. The main
10 conclusion has been that the null hypothesis of equality between the CVM value of a forest and
11 the value predicted from inference cannot be rejected. The result suggests that, if it could be
12 further replicated, the combination of CVM and CE (base and marginal values) would constitute
13 a reliable alternative to valuing relatively similar environmental goods.

14
15 Future work could test whether assuming other than a linear relation between attribute level and
16 value, or using discrete instead of marginal values of attributes, could improve the accuracy of
17 the inference. Additionally, using CE only, without relaying in CVM, could also be considered.

18

19 **References**

20 Adamowicz, W.L., Louviere, J.J. and Williams, M. 1994. Combining stated and revealed
21 preference methods for valuing environmental amenities. *Journal of Environmental Economics*
22 *and Management*, **263**: 271-292.

23

- 1 Adamowicz, W.L., Swait, J., Boxall, P., Louviere, J. and Williams, M. 1996. Perceptions versus
2 objective measurement of environmental quality in combined revealed and stated preference
3 models of environmental valuation. *Journal of Environmental Economics and Management*, **32**:
4 65-84.
- 5
- 6 Adamowicz, W.L., Boxall, P.C., Williams, M. and Louviere, J.J. 1998. Stated preference
7 approaches for measuring passive use values: choice experiments and contingent valuation.
8 *American Journal of Agricultural Economics*, **80**: 65-75.
- 9
- 10 Amemiya, T. 1984. Qualitative response models: a survey. *Journal of Economic Literature*, **21**:
11 1483-1536.
- 12
- 13 Arrow, K., Solow, R., Portney, P., Leamer, E., Radner, R. Schuman, H. 1993. Report of the
14 NOAA Panel on contingent valuation. *Federal Register*, **58**: 4602-4614.
- 15
- 16 Azqueta, D. and Touza-Montero, J. 2000. Benefit transfer: the recreational value of three natural
17 parks in Galicia. *In Spatial Economics and Ecosystems. The interaction between Economics and*
18 *the Natural Environment. Edited by WIT Press, Southampton, U.K. pp: 257-283.*
- 19
- 20 Ben-Akiva, M. and Bolduc, D. 1987. Approaches to model transferability and updating: the
21 combined transfer estimator *Transportation Research Record*, **1139**: 1-7.
- 22

- 1 Bergland, O. 1997. Valuation of landscape elements using a contingent choice method. *In*
2 proceedings of the 8th Conference of the European Association of Environmental and Resource
3 Economists, Tilburg, Netherlands.
- 4
- 5 Bergland, O., Magnussen, K. and Navrud, S. 1995. Benefit transfer: testing for accuracy and
6 reliability, Discussion Paper, #D-03/1995, Department of Economics and Social Sciences,
7 University of Norway.
- 8
- 9 Bergstrom, J.C. and De Civita, P. 1999: Status of benefit transfer in the United States and
10 Canada: a review *Canadian Journal of Agricultural Economics*, **47**: 79-87.
- 11
- 12 Boxall, P., Adamowicz, W.L., Williams, M., Swait, J. and Louviere, J.J. 1996. A comparison of
13 stated preference approaches to the measurement of environmental values. *Ecological*
14 *Economics*, **18**: 243-253.
- 15
- 16 Boxall, P. and Macnab, B. 2000. Exploring the preferences of wildlife recreationists for features
17 of boreal forest management: a choice experiment approach. *Canadian Journal of Forest*
18 *Research*, **30**: 1931–1941.
- 19
- 20 Boyle, K.J. and Bergstrom, J.C. 1992. Benefit transfer studies: myths, pragmatism and idealism.
21 *Water Resources Research*, **28**: 657-663.
- 22
- 23 Brouwer, R. 2000. Environmental value transfer: state of the art and future prospects *Ecological*
24 *Economics*, **32**: 137-152.

1
2 Christie, M and Azevedo, C. 2002. Testing the consistency in benefit estimates across contingent
3 valuation and choice experiments: a multiple policy option application. In proceedings of the 2nd
4 World Congress of Environmental and Resource Economists, June, Monterrey, California.

5 Departament de Medi Ambient 1996. Les emissions a l'atmosfera a Catalunya. 1996 Una
6 aproximació quantitativa. *Quaderns de medi ambient*, 5. Generalitat de Catalunya, Barcelona.

7
8 Desvousges, W.H., Naughton, M.C. and Parsons, G.R. 1992. Benefit transfer: conceptual
9 problems in estimating water quality benefits using existing studies. *Water Resources Research*,
10 **28**: 675-683.

11
12 Diamond, Peter A. and Hausman, Jerry A. 1994. Contingent valuation: is some number better
13 than no number? *Journal of Economic Perspectives*, **8**: 45-64.

14
15 Economics for the Environment Consultancy (EFTEC) 2000. Guidance on using stated
16 preference techniques for the economic valuation of non-market effects. Report to the
17 Department of the Environment, Transport and the Regions (DETR), UK.

18
19 Efron, B. and Tibshirani, R.J. 1993. *An Introduction to the Bootstrap*. Chapman and Hall, New
20 York.

21
22 Eom, Y.S. 1994. Pesticide residue risk and food safety valuation: a random utility approach.
23 *American Journal of Agricultural Economics*, **76**: 760-771.

24

1 Galbraith, R. and Hensher, D.A. 1982. Intra-Metropolitan transferability of mode choice models.
2 *Journal of Transport Economics and Policy*, **16**: 7-29.

3
4 Haener, M., Boxall, P.C. and Adamowicz, W.L. 2001. Modeling recreation site choice: Do
5 hypothetical choices reflect actual behavior? *American Journal of Agricultural Economics*, **83**:
6 629-642.

7
8 Hanemann, W.M. 1984. Welfare evaluations in contingent valuation experiments with discrete
9 responses. *American Journal of Agricultural Economics*, **66**: 332-341.

10
11 Hanemann, W.M. and Kanninen, B. K. 1999. The statistical analysis of discrete-response CV
12 data. In *Valuing Environmental Preferences: Theory and Practice of the Contingent Valuation*
13 *Method in the US, EC, and Developing Countries*, Oxford: Oxford University Press; chapter 11.

14
15 Hanley, N., MacMillan, D., Wright, R.E., Bullock, C., Simpson, I., Parsisson, D. and Crabtree, B.
16 1998a. Contingent valuation versus choice experiments: estimating the benefits of
17 environmentally sensitive areas in Scotland. *Journal of Agricultural Economics*, **49**: 1-15.

18
19 Hanley, N., Wright, R.E. and Adamowicz, W. 1998b. Using choice experiments to value the
20 environment. *Environmental and Resource Economics*, **11**: 413-428.

21
22 Hanley, N., Mourato, S. and Wright, R. 2001. Choice modelling approaches: a superior
23 alternative for environmental valuation? *Journal of Economic Surveys*, **15**: 433-460.

24

- 1 Kask, S.B. and Shogren, J.F. 1994. Benefit transfer protocol for long-term health risk valuation: a
2 case of surface water contamination. *Water Resources Research*, **30**: 2813-2823.
3
- 4 Kirchoff, S., Colby, B.G. and LaFrance, J.F. 1997. Evaluation the performance of benefit
5 transfer: an empirical inquiry. *Journal of Environmental Economics and Management*, **33**: 75-93.
6
- 7 Koppelman, F.S., Kuah, G. and Wilmot, C.G. 1985. Transfer model updating with disaggregate
8 data. *Transportation Research Record*, **1037**: 102-107.
9
- 10 Krinsky, I. and Robb, L.A. 1986. On approximating the statistical properties of elasticities. *The*
11 *Review of Economics and Statistics*, **68**: 715-719.
12
- 13 Loomis, J.B. 1992. The evolution of a more rigorous approach to benefit transfer: benefit
14 function transfer. *Water Resources Research*, **28**: 701-705.
15
- 16 Louviere, J.J. 1988. Analysing individual decision making: metric conjoint analysis. Sage
17 university series on quantitative applications in the social sciences, n° 67. Newbury Park, CA:
18 Sage Publications, Inc.
19
- 20 Louviere, J.J. 1991. Experimental choice analysis: introduction and overview. *Journal of*
21 *Business Research*, **23**: 291-297.
22
- 23 Louviere, J.J., Hensher, D.A., and Swait, J. 2000. Stated choice methods: analysis and
24 applications. Cambridge University Press.






- 1
- 2 McFadden, D. 1974 Conditional logit analysis of qualitative choice behaviour. *In* *Frontiers in*
- 3 *Econometrics*, New York: Academic Press, pp. 105-142.
- 4 Ministerio de Medio Ambiente 1996. Segundo Inventario Nacional, 1986-1996. Dirección
- 5 General de Conservación de la Naturaleza, Ministerio de Medio Ambiente, Madrid.
- 6
- 7 Mitchell, R. C. and Carson, R. T. 1989. Using surveys to value public goods. The contingent
- 8 valuation method. *Resources For the Future*, Washington, DC.
- 9
- 10 Mogas, J., Riera, P. and Bennett, J. 2002. A comparison of contingent valuation and choice
- 11 modelling: estimating the environmental values of Catalanian Forest. *In* Working Paper series,
- 12 nº, 1. National Centre for Development Studies Asia Pacific School of Economics and
- 13 Management. Australian National University, Australia.
- 14
- 15 Morrison, M., Bennett, J., Blamey, R. and Louviere, J. 1998. Choice modelling and test of benefit
- 16 transfer. Choice Modelling Research Report 8, University College, The university of New South
- 17 Wales, Canberra, Australia.
- 18
- 19 Morrison, M. and Bennett, J. 2000. Choice modelling, non-use values and benefit transfer.
- 20 *Economics Analysis and Policy*, **30**: 13-32.
- 21
- 22 Morrison, M., Bennett, J., Blamey, R. and Louviere, J. 2002. Choice modeling and tests of
- 23 benefit transfer, *American Journal of Agricultural Economics*, **84**: 161-170
- 24

- 1 Opaluch, J.J. Swallow, S.K., Weaver, T., Wessells, W. and Wichelns, D. 1993. Evaluating
2 impacts from noxious facilities: including public preferences in current siting mechanisms.
3 *Journal of Environmental Economics and Management*, **24**: 41-59
- 4 Organisation for Economic Co-operation and Development (OECD). 1993. Project and policy
5 appraisal: Integrating economics and environment. Group on Economic and Environmental
6 Policy Integration, Environment Directorate, Environmental Policy Committee,
7 ENV/EPOC/GEEI(93) 14, Paris, France.
- 8
- 9 Poe, G., Welsh, M. and Champ, P. 1997. Measuring the difference in mean willingness to pay
10 when dichotomous choice contingent valuation responses are not independent. *Land Economics*,
11 **73**: 255-267.
- 12
- 13 Rolfe, J. and Bennett, J. 1996. Valuing international rainforests: a choice modelling approach. *In*
14 the proceedings of the *40th Annual Conference of the Australian Society of Agricultural*
15 *Economics*, Melbourne.
- 16
- 17 Rolfe, J. Bennett, J. and Louviere, J. 2002. Stated values and reminders of substitute goods:
18 Testing for framing effects with choice modelling. *The Australian Journal of Agricultural and*
19 *Resource Economics*, **46**: 1-20.
- 20
- 21 Rosenberger, R. S. and Loomis, J. B. 2001. Benefit transfer of outdoor recreation use values: A
22 technical document supporting the Forest Service Strategic Plan (2000 revision). Gen. Tech. Rep.
23 RMRS-GTR-72. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky
24 Mountain Research Station.

- 1
- 2 Ruijgrok, E.C.M. 2001. Transferring economic values on the basis of an ecological
3 classification of nature. *Ecological Economics*, **39**: 399-408.
- 4
- 5 Smith, V.K. 1992. On separating defensible benefit transfer from “Smoke and Mirrors”. *Water*
6 *Resources Research*, **28**:685-694.
- 7
- 8 Swait, J., Louviere, J. 1993. The role of the scale parameter in the estimation and comparison of
9 multinomial logit models. *Journal of Marketing Research*, **30**: 305-314.
- 10
- 11 Thurstone, L. 1927. A law of comparative judgment. *Psychological Review*, **34**: 273-286.
- 12
- 13 Veall, Michael, R., and Zimmermann, Klaus F. 1996. Pseudo-R² measures for some common
14 limited dependent variable models. *Journal of Economic Surveys*, **10**: 241-59.
- 15
- 16 Xu, X. and Adamowicz, W. 1997. Modelling non-timber benefits via benefit transfer. Network of
17 Centres of Excellence, Project Report SES-1, University of Alberta, Edmonton, Alberta, Canada.

Example of a pair of afforestation alternatives used in the CE exercise

Of these two ways of afforesting 10% of the area of Catalonia [SHOW CARD AND EXPLAIN THE MEANING OF THE LEVELS AND ATTRIBUTES]

FOREST UTILITY	AFFORESTATION A	AFFORESTATION B
RECREATIONAL ACTIVITIES ALLOWED	 <ul style="list-style-type: none"> • DRIVING BY CARS IN DESIGNATED TRAILS 	 <ul style="list-style-type: none"> • DRIVING BY CARS IN DESIGNATED TRAILS  <ul style="list-style-type: none"> • PICNICKING
GAS CO₂ SEQUESTERED PER YEAR (Pollution produced by....)	 <p>CITY OF 400.000 INHABITANTS</p>	 <p>CITY OF 500.000 INHABITANTS</p>
NEW FOREST IN...	<p>Soil eroded A LITTLE (Unproductive in 500 years)</p>	<p>Soil eroded A LITTLE (Unproductive in 500 years)</p>
<u>ECONOMIC COST</u>		
ANNUAL CONTRIBUTION	18 euros/year	24 euros/year

¿Would you choose A, B, or neither?[**ALLOW INDIFFERENCE**]





- AFFORESTATION A.....1
- AFFORESTATION B.....2
- NEITHER.....3
- INDIFERENT.....4
- DOESN'T KNOW..... 8
- DOESN'T ANSWER.....9

Appendix B.

Example of the question used in the CVM corresponding to the FOREST B

The afforestation that the administration is thinking about carrying out is the afforestation corresponding to the card [SHOW CARD] ,i.e., they would do a afforestation with these effects [EXPLAIN THE MEANING OF THE LEVELS AND ATTRIBUTES OF THE CARD]

FOREST B

FOREST UTILITY	
RECREATIONAL ACTIVITIES ALLOWED	 <ul style="list-style-type: none"> • DRIVING BY CARS IN DESIGNATED TRAILS
	 <ul style="list-style-type: none"> • PICNICKING
	 <ul style="list-style-type: none"> • PICKING MUSHROOMS
GAS CO₂ SEQUESTERED PER YEAR (Pollution produced by....	 CITY OF 400.000 INHABITANTS
NEW FOREST IN.....	Soil eroded A LITTLE (Unproductive in 500 years)
ECONOMIC COST	
ANNUAL CONTRIBUTION	36 euros/year

¿Would you accept this afforestation?

- YES..... 1
- NO..... 2
- DOESN'T KNOW..... 8
- DOESN'T ANSWER.... 9